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Positive effects of tree species diversity on litterfall quantity and quality along a secondary successional chronosequence in a subtropical forest

Huang, Yuanyuan ; Ma, Yinlei ; Zhao, Ke ; Niklaus, Pascal A ; Schmid, Bernhard ; He, Jin-Sheng

Abstract: Aims Litterfall, as an important link between aboveground and belowground processes, plays a key role in forest ecosystems. Here, we test for effects of tree species richness on litter production and litter quality in subtropical forest. The study further encompasses a factorial gradient of secondary succession that resulted from human exploitation. Given that a large percentage of subtropical forests are in secondary successional stages, understanding the role of biodiversity on forest re-growth after disturbance appears critical. Methods From January 2009 to December 2014, we monitored forest litterfall in 27 Comparative Study Plots that spanned a gradient of tree species richness (3–20 species) and secondary successional ages (20 to 120 years) in Gutianshan Natural Nature Reserve, Zhejiang Province, China. The experiment is part of the biodiversity–ecosystem functioning research platform ‘BEF-China’. Tree litterfall was collected in monthly intervals using litter traps. Samples were separated into leaf and non-leaf components. Leaf litter was further sorted into dominant and other species. Community level monthly leaf litter C and N contents were analysed through a full year. General linear mixed-effects models were applied to test for effects of tree species richness and successional age on litter quantity and leaf litter C/N. Important Findings Litterfall increased with species richness among and within successional age and this effect was consistent across years. Successionally older stands had higher litterfall and this effect was related to increased tree species richness. However, species richness did not change the intra- and inter-annual temporal stability of litterfall. Increasing tree species richness increased leaf litter quality (decreased C/N), while successional age had no effect. Our study indicates that more diverse forest stands produce more leaf litter and that this litter has higher N concentrations, which could promote forest growth through accelerated nutrient re-cycling.

DOI: <https://doi.org/10.1093/jpe/rtw115>

Posted at the Zurich Open Repository and Archive, University of Zurich

ZORA URL: <https://doi.org/10.5167/uzh-147985>

Journal Article

Accepted Version

Originally published at:

Huang, Yuanyuan; Ma, Yinlei; Zhao, Ke; Niklaus, Pascal A; Schmid, Bernhard; He, Jin-Sheng (2017). Positive effects of tree species diversity on litterfall quantity and quality along a secondary successional chronosequence in a subtropical forest. *Journal of Plant Ecology*, 10(1):28-35.

DOI: <https://doi.org/10.1093/jpe/rtw115>

1 **Positive effects of tree species diversity on litterfall quantity**
2 **and quality along a secondary successional chronosequence**
3 **in a subtropical forest**

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14 **Abstract**

15 **Aims**

16 Litterfall, as an important link between aboveground and belowground processes,
17 plays a key role in forest ecosystems. Here, we test for effects of tree species richness
18 on litter production and litter quality in subtropical forest. The study further
19 encompasses a factorial gradient of secondary succession that resulted from human
20 exploitation. Given that a large percentage of subtropical forests are in secondary
21 successional stages, understanding the role of biodiversity on forest re-growth after
22 disturbance appears critical.

23 **Methods**

24 From January 2009 to December 2014, we monitored forest litterfall in 27
25 Comparative Study Plots (CSPs) that spanned a gradient of tree species richness (3 –
26 20 species) and secondary successional ages (~ 20 – 120 years) in Gutianshan Natural
27 Nature Reserve, Zhejiang Province, China. The experiment is part of the biodiversity–
28 ecosystem functioning research platform “BEF-China”. Tree litterfall was collected in
29 monthly intervals using litter traps. Samples were separated into leaf and non-leaf
30 components. Leaf litter was further sorted into dominant and other species.
31 Community level monthly leaf litter C and N contents were analysed through a full
32 year. General linear mixed-effects models were applied to test for effects of tree
33 species richness and successional age on litter quantity and leaf litter C/N.

34 **Important Findings**

35 Litterfall increased with species richness among and within successional age and this
36 effect was consistent across years. Successionally older stands had higher litterfall,
37 and this effect was related to increased tree species richness. However, species

38 richness did not change the intra and inter-annual temporal stability of litterfall.

39 Increasing tree species richness increased leaf litter quality (decreased C/N), while

40 successional age had no effect.

41 Our study indicates that more diverse forest stands produce more leaf litter and that

42 this litter has higher N concentrations, which could promote forest growth through

43 accelerated nutrient recycling.

44 **Keywords:** species richness, BEF-China, litterfall, leaf litter C/N, subtropical forest,

45 secondary succession, structural equation models

46 **Introduction**

47 Above- and belowground carbon dynamics are linked through litterfall, which thus is
48 a key determinant of forest ecosystem functioning (Sayer *et al.*, 2010). Positive
49 effects of tree species richness on forest productivity have been found in several
50 recent studies (Barrufol *et al.*, 2013, Morin *et al.*, 2011, Zhang *et al.*, 2012, Chen *et*
51 *al.*, 2016, Chisholm *et al.*, 2013, LaManna *et al.*, 2016). However, diversity effects on
52 litter production and litter quality have received less attention.

53 Previous studies have found a positive correlation between net primary
54 productivity (NPP) and litterfall, with aboveground litter fluxes accounting for
55 approximately one third of NPP (Clark *et al.*, 2001a, Gower *et al.*, 1997, He *et al.*,
56 2012, Nouvellon *et al.*, 2012). Increased productivity in more diverse forest (Baruffol
57 *et al.*, 2013) raises the possibility that litter production increases with species richness,
58 which might lead to faster nutrient cycling. However, nutrient losses and re-cycling
59 rates also depend on leaf litter quality (Li *et al.*, in press). One of the important
60 indicators of leaf litter quality is its carbon to nitrogen ratio (C/N), which depends on
61 leaf nutrient status and nutrient resorption and reflects the nutrient-use strategies of
62 species and individuals (Vitousek, 1984). At the community level, the quality of
63 forest leaf litter will likely depend on the species composition of tree stands. Apart
64 from that, leaf litter quality of the same species may also change with the diversity of
65 the community in which they grow.

66 Positive effects of tree species richness on litterfall have been found in a tropical
67 forest biodiversity experiment (Scherer-Lorenzen *et al.*, 2007), but only at low
68 diversity levels (3-species mixtures compared with monocultures), with no additional
69 effect of a higher species richness level (6-species mixtures). In the same study,
70 effects on leaf litter C and N content were highly species-specific (Scherer-Lorenzen

71 *et al.*, 2007). A number of studies have compared litterfall in monocultures to litterfall
72 in natural forests (e.g. Yang *et al.*, 2004). However, natural forests generally have a
73 more complex stand structure and differ in demographic dynamics, so that it is
74 difficult to infer effects of tree species richness from a comparison with planted
75 monocultures (Coursolle *et al.*, 2012, He *et al.*, 2012, Yang *et al.*, 2011).

76 With the rapid increase in the human exploitation of natural resources, an
77 increasing number of forests are in secondary successional stages. At later
78 successional stages, tree growth slows down compared to younger stands (Chi *et al.*,
79 in press). Older stands generally are characterized by a higher number of canopy
80 species, fewer shade-tolerant species, higher standing biomass (Baruffol *et al.*, 2013)
81 and a higher investment into defense against herbivores and pathogens leading to
82 lower leaf N content (Bruehlheide *et al.*, 2011, Kröber *et al.*, 2012). For these reasons,
83 effects of tree species richness on litterfall may depend on stand age.

84 To date, long-term investigations of the relationship between biodiversity and
85 litterfall quantity and quality in natural forests with complex structure are scarce. It
86 thus remains unclear how species richness contributes to the restoration of ecosystem
87 processes and services through secondary forest succession.

88 Here, we measured tree litter production for six years in plots spanning largely
89 independent gradients of tree species richness and forest successional age in species-
90 rich subtropical forest. We hypothesized that (1) litterfall increases with species
91 richness and that this effect increases with successional age; (2) tree species richness
92 decreases leaf litter C/N, i.e. improves leaf litter quality, and more strongly so in later
93 successional stages.

94

95 **Materials and Methods**

96 **Study site**

97 The present study was carried out in Gutianshan National Nature Reserve in the
98 western part of Zhejiang Province, China (29° 8' 18" – 29° 17' 29" N, 118° 2' 14" –
99 118° 11' 12" E). This region has a typical subtropical monsoon climate with an annual
100 average temperature of 15 °C and annual average precipitation of *ca.* 2000 mm. The
101 bedrock is comprised of granite and gneiss. Sandy-loamy and silty-loamy acidic
102 Cambisols with pH ranging from 4 to 5 are the predominant soil type (Geißler *et al.*,
103 2010).

104 Deforestation at the study site occurred during different periods since the 1950s,
105 resulting in a patch structure with respect to successional age. Forest patches also vary
106 in species richness, presumably due to variation in seed rain, natural recruitment and
107 environmental conditions. In 2009, we established 27 plots with a size of 30 x 30 m
108 each, called Comparative Study Plots (CSPs). These plots span factorial gradients in
109 tree diversity and successional ages (Bruehlheide *et al.*, 2011, Table S1
110 (Supplementary Data)). Stand age was defined as the age of the fifth-largest tree in a
111 plot, with age determined from a stem core (Bruehlheide *et al.*, 2011). Because age is
112 not a precise metric, we assigned plots to three age classes (young forest: 20 – 50;
113 medium forest: 50 – 80; old forest: > 80 years old).

114 In 2008, an inventory was conducted to assess tree species composition of each
115 plot (Baruffol *et al.*, 2013). Canopy trees, defined here by a diameter at breast height
116 of at least 10 cm, comprised 1523 individuals belonging to 66 species, 49 genera, and
117 29 families. In the present study, we use canopy tree species richness as metric of
118 biodiversity for all analysis. The reason for this choice was that the litter traps we
119 used were installed 1.5 m above ground and therefore mainly collected litter from
120 canopy trees. Similar to successional age, plot-level tree species richness was

121 categorized into three classes (low: 3 – 8, medium: 9 – 13, high: 14 – 20 species) to
122 reflect the deliberate selection of plots belonging to these three richness categories.

123 **Litterfall**

124 In December 2008, four litter traps were set up in the corners of the central 10 × 10 m
125 quadrat plus one in the middle of each plot. A nylon net (1 mm mesh) with a
126 horizontal trapping area of 0.75 × 0.75 m was placed over a PVC frame 1.5 m above
127 the ground.

128 Litterfall was collected monthly from January 2009 to December 2014. In
129 December 2010 and July 2011, litter could not be collected because of heavy snow
130 and rainstorm. Litterfall of these months was collected together with litter from the
131 next month, and this amount was partitioned among the respective months based on
132 the litter distribution in the other years. In 2010 and 2011, litter was first separated
133 into leaf and non-leaf litter. The leaf litter was then sorted into the dominant species
134 (*Castanopsis eyrei*, *Schima superba*, *Pinus massoniana*, *Cyclobalanopsis glauca*,
135 *Quercus serrata* var. *brevipetiolata*, *Lithocarpus glaber*) and other evergreen and
136 deciduous litter. The number of other evergreen and deciduous litter species per trap
137 was also recorded from January to March of year 2010 and from April to December
138 of year 2011. Non-leaf litter included fine branches (≤ 2.5 cm in diameter), bark,
139 reproductive structures, animal detritus and other unidentified fine litter.

140 Litter was weighed after oven-drying at 80°C for 24 h. Leaf litter samples of the
141 year 2010 were pooled by plot, ground using a ball mill (NM200, Retsch, Haan,
142 Germany), and C and N concentrations determined by dry combustion (2400 II CHN
143 elemental analyzer, Perkin-Elmer, USA).

144 **Statistical analyses**

145 The effects of year (1 – 6), month (1 – 12), successional age (1 – 3), species richness
146 (3 – 20) and their interactions on the litterfall and leaf litter quality were analyzed
147 with linear mixed-effects models using ASReml-R (Butler *et al.*, 2007). All fixed-
148 effects terms in the model were fitted sequentially. Litterfall amounts were square
149 root-transformed prior to analysis to meet the requirements of normal distribution and
150 homoscedasticity of residuals.

151 For yearly litterfall data, the fixed-effects terms were fitted in this sequence:
152 YEAR + div + lin(age) + AGE + YEAR × div + YEAR × lin(age) + YEAR × AGE +
153 div × lin(age) + div × AGE. The random-effects terms used were plot, modeling the
154 random sampling of plots, and the interaction plot×YEAR, which corresponded to the
155 residual. Interactions are denoted by a × operator. Capitalized terms YEAR (6 levels)
156 and AGE (3 levels) are factors, while lin(age) is a continuous integer variable, i.e. the
157 linear contrasts of factor AGE. Similarly, div is a continuous integer variable and
158 stands for species richness.

159 Monthly litterfall data were analyzed with a mixed-effects model with plot, plot ×
160 MONTH and plot × YEAR as random effects and factor MONTH (12 levels), species
161 richness, successional age and their interactions as fixed terms. Leaf C/N was
162 analyzed similarly excluding terms containing YEAR.

163 Because plot species richness was positively correlated with successional age ($r =$
164 0.64 , $P < 0.01$), we reversed the order of tree species richness and successional age to
165 investigate the degree of confounding between their effects (Baruffol *et al.*, 2013;
166 Schmid *et al.*, in press). Species richness effects fitted before successional age
167 indicated the overall species richness effect, whereas richness fitted after successional

168 age indicated effects of species richness after adjusting for successional age (i.e.
169 effects of species richness within successional age).

170

171 **Results**

172 **Environmental effects**

173 Site conditions (Table S1), including elevation, slope, aspect, slope inclination, soil
174 pH, soil moisture and soil total C and N content did not correlate significantly with
175 tree litterfall or leaf litter C/N (Pearson's product moment correlations, $P > 0.05$).

176 **Litterfall quantity**

177 Annual litterfall varied among years and plots, ranging from 2.6 Mg ha⁻¹ yr⁻¹ to 7.9
178 Mg ha⁻¹ yr⁻¹, with a mean value of 5.4 Mg ha⁻¹ yr⁻¹.

179 Canopy tree species richness significantly positively affected yearly total litterfall
180 (Fig. 2a, Table 1a, $P < 0.001$ in mixed-effects model when div was fitted before
181 $\ln(\text{age})$ and AGE; $P < 0.05$ when div was fitted after $\ln(\text{age})$ and AGE). The positive
182 effect of species richness was also observed in a structural equation model (Fig. S1
183 (Supplementary Data)). The positive effect of species richness was similar at different
184 successional ages (Fig. 2a, Table 1a, $P = 0.64$ for $\text{div} \times \ln(\text{age})$). Species richness
185 effects were independent of year (Table 1a, $P > 0.05$ for $\text{YEAR} \times \text{div}$).

186 Yearly litterfall increased with forest stand age if influences of species richness
187 were ignored (Fig. 2a, Table 1a, $P < 0.01$ for $\ln(\text{age})$ fitted before div). However, the
188 main effect of successional age became statistically insignificant when it was adjusted
189 for species richness (Table 1a, $P > 0.1$ for $\ln(\text{age})$ fitted after div). The notion that
190 age effects might have been mediated by species richness changes was supported by
191 structural equation modelling (Fig. S1, no significant direct path from successional

age to litterfall). Litterfall gradually increased with year (Fig. 1, anova with $\text{lin}(\text{year})$ as continuous variable, $F_{1,134} = 18.27$, $P < 0.01$), and this effect was more pronounced in medium or old forest (Fig. 1).

The analysis of monthly litterfall data revealed a dependence of species richness effects on season (Table 1b, $P < 0.01$ for $\text{MONTH} \times \text{div}$; Fig. 3). Different litterfall components had different monthly dynamics (Fig. 3). Leaf litterfall showed a bimodal temporal trend, whereas non-leaf litterfall did not. Positive species richness effects were found for total and leaf litterfall only in months with high litter production (Apr. – May, Oct. – Nov.; Fig. 3a, b).

Neither the intra-annual (seasonal) nor the inter-annual stability of litterfall production depended on species richness (Fig. 3, Fig. S2 (Supplementary Data)).

Litter species composition

Higher species numbers were found in the traps in more species-rich plots, especially in months with high litterfall (Fig. 5). As for litterfall, litter species richness followed a bimodal temporal pattern for evergreen species, but a unimodal pattern for deciduous species (Fig. 5). In general, litterfall from the highly productive dominant species *Schima superba* and *Castanopsis eyrei*, as well as other evergreen and other deciduous species, was higher in species-rich plots and in later successional stages (Fig. 6).

Leaf litter C/N

Leaf litter C/N, a proxy for litter quality, averaged 46.4 ± 1.1 g C (g N)⁻¹ ($50.2\% \pm 14\%$ C; $1.1\% \pm 3\%$ N). Total leaf litter C return was 1720 ± 123 kg C ha⁻¹ yr⁻¹ and N return 38.6 ± 2.8 kg N ha⁻¹ yr⁻¹. Leaf litter C/N decreased significantly with increasing tree species richness (Fig. 2b, Table 1c, $P < 0.05$ with div fitted before $\text{lin}(\text{age})$ and

216 AGE; $P < 0.05$ with div fitted after $\ln(\text{age})$ and AGE). Leaf C/N did not depend on
217 successional age (Fig. 2b, Table 1c, $P > 0.05$ for $\ln(\text{age})$ and AGE fitted before or
218 after $\log(\text{div})$).

219

220 **Discussion**

221 Our results showed a strong positive effect of species richness on both litterfall
222 amounts and litter quality (Tables 1a, b). This finding parallels strong positive,
223 density-mediated tree diversity effects on stand total basal area and growth in the
224 same plots (Baruffol *et al.* 2013). Baruffol *et al.* (2013) argued that the larger number
225 of tree individuals found in more diverse plots possibly resulted from
226 complementarity among species, i.e. that more complementarity among species
227 reduced competition among heterospecific trees and allowed for denser stands at
228 higher species richness. Belowground benefits from complementarity may also have
229 contributed to higher total leaf production (Bessler *et al.*, 2009; Bu *et al.*, in press;
230 Sun *et al.*, in press). Interestingly, positive effects of species richness on leaf area
231 were also found in a designed experiment with constant tree density (Peng *et al.*, in
232 press). The increase in litterfall with diversity may thus also be caused by higher leaf
233 production of individual trees in more diverse forest stands (Clark *et al.*, 2001a).
234 Overall, this suggests that more diverse forest stands produce more leaf litter, and that
235 this effect can but must not necessarily be mediated by an increased density of tree
236 individuals, but also by increasing leaf production.

237 The higher leaf litter nitrogen contents in more diverse plots may indicate an
238 improved supply of trees with nitrogen and/or a lower nutrient resorption efficiency.
239 In forests, nutrient concentrations of newly fallen leaf litter often correlate positively
240 with nutrient concentrations of fresh leaves (Aponte *et al.*, 2013). An additional

possible explanation for the observed effects in our study is that the fraction of deciduous broadleaf trees increased with species richness (Fig. 6a). These species generally have higher green fresh leaf N concentrations than evergreen or conifer species (Han *et al.*, 2005, Kröber *et al.*, 2012, McGroddy *et al.*, 2004). It is conceivable that this change in species composition with increasing species richness was part of the mechanism underlying the positive species richness effects on litterfall N fluxes. Because the quality of litterfall is a major controller of litter decomposition and nutrient return rates (Manzoni *et al.*, 2008, Meier *et al.*, 2008), the lower initial litter C/N ratio in more species-rich plots in our study could accelerate decomposition and nitrogen mineralization. Tree species diversity might thus promote productivity by increasing nutrient availability.

The species richness effects we found were independent of successional age. Our hypotheses, predicting a stronger biodiversity effect in later successional stages were rejected. This suggests that even in young forests tree species diversity already plays an important role. Higher litterfall in old successional forests has been attributed to higher standing leaf canopy as well as decreased physiological function of older trees (Drake *et al.*, 2011). Our analysis suggests that increased species richness may be a further factor affecting increased litter fall in older secondary forests, which is consistent with the effects on woody growth pattern (Baruffol *et al.*, 2013; Fig. S1, Tables 1a, b).

Our findings contrast with those of other studies in which no significant effects of tree diversity on litter production and litter N content were found (Scherer-Lorenzen *et al.*, 2007). However, in those previous studies lower diversity levels were compared (species richness levels 1, 3, 6) whereas in our study the range of species richness values was considerably higher (3 – 20 species).

266 In our study we did not find a significant effect of successional age on the leaf
267 litter C/N (Table 1c). This does not support the assumption that there should be a
268 change in resource-use strategy from high nutrient acquisition to nutrient retention
269 along forest succession. It also does not support the assumption that trees should
270 increasingly allocate more energy to defense which may decrease leaf litter quality
271 (Kröber *et al.*, 2012). However, regarding the first assumption, nutrient resorption
272 efficiency may also be higher in earlier successional stage, serving the higher demand
273 for nutrients in faster-growing trees (Yuan *et al.*, 2010). Regarding the second
274 assumption, former research in the same study plots has shown that while green leaf
275 physical resistance increased, chemical defense traits, such as tannin and phenolics,
276 decreased with forest successional age, such that the litter decomposition rates
277 remained stable along secondary succession (Eichenberg *et al.*, 2014). Positive and
278 negative effects may have acted together and thus caused the balanced leaf litter C/N
279 across successional age in our study plots.

280 In conclusion, tree species diversity rather than forest successional age seemed to
281 play the major role in affecting leaf litter quantity and quality. Higher litterfall
282 production and better leaf litter quality in more diverse forest stands could promote
283 higher soil microbial and fauna diversity and create more favorable conditions for
284 decomposition and nutrient release, thus stimulating increased tree growth.

285

286

287 **Supplementary Data**

288 We present the following supplementary data:

289 Table S1: Topography and soil information for the 27 Comparative Study Plots.

290 Figure S1: Structural equation model.

291 Figure S2: Litterfall stability as function of species richness.

292

293 **Funding**

294 This study was supported by the EU 7th FP Project IDP BRIDGES (Grant number

295 608422) and the National Basic Research Program of China (grant number

296 2014CB954004).

297

298 **Acknowledgements**

299 We are grateful to Limei Guo for the help of C and N analysis, and we thank Yan

300 Chen, Zaigen Jiang, Zhenglin Lai for the help during field work. We would also like

301 to thank the reviewers for very helpful suggestions on earlier versions of this paper.

302

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408

409 Table 1: results of three alternative sequential mixed-effects models for yearly,
 410 monthly litter production and leaf litter C/N

411 (a) yearly litterfall data from 2009 – 2014

model 1			model 2		
Fixed terms	<i>F</i> -value		Fixed terms	<i>F</i> -value	
YEAR	F(5,115)=26.51	***	YEAR	F(5,115)=26.51	***
div	F(1,21)=11.47	***	lin(age)	F(1,21)=9.82	**
lin(age)	F(1,21)=1.55	n.s.	AGE	F(1,21)=0.12	n.s.
AGE	F(1,21)=1.42	n.s.	div	F(1,21)=4.50	*
YEAR×div	F(5,115)=2.14	.	YEAR×lin(age)	F(5,115)=3.71	**
YEAR×lin(age)	F(5,115)=2.26	.	YEAR×AGE	F(5,115)=2.63	*
YEAR×AGE	F(5,115)=3.61	**	YEAR×div	F(5,115)=1.67	n.s.
div×lin(age)	F(1,21)=0.00	n.s.	lin(age)×div	F(1,21)=0.00	n.s.
div×AGE	F(1,21)=0.23	n.s.	AGE×div	F(1,21)=0.23	n.s.
Random terms	Variance component		Standard error of variance component		
plot	3.0437		1.0179		
residuals	1.5172		0.2001		

412 (b) monthly litterfall data from 2009 – 2014

model 1			model 2		
Fixed terms	<i>F</i> -value		Fixed terms	<i>F</i> -value	
MONTH	F(11,253)=81.68	***	MONTH	F(11,253)=81.68	***
div	F(1,21)=9.91	**	lin(age)	F(1,21)=7.73	*
lin(age)	F(1,21)=0.97	n.s.	AGE	F(1,21)=0.02	n.s.
AGE	F(1,21)=0.96	n.s.	div	F(1,21)=4.08	.
MONTH×div	F(11,253)=1.88	*	MONTH×lin(age)	F(11,253)=1.84	*
MONTH×lin(age)	F(11,253)=1.26	n.s.	MONTH×AGE	F(11,253)=0.51	n.s.
MONTH×AGE	F(11,254)=1.07	n.s.	MONTH×div	F(11,253)=1.86	*
lin(age)×div	F(1,21)=0.00	n.s.	lin(age)×div	F(1,21)=0.00	n.s.
AGE×div	F(1,21)=0.73	n.s.	AGE×div	F(1,21)=0.73	n.s.
Random terms	Variance component		Standard error of variance component		
plot	0.2071		0.0905		
plot × YEAR	0.2526		0.0359		
plot × MONTH	0.4306		0.0459		
residuals	2.5165		0.0372		

413

414 (c) monthly leaf litter C/N data of year 2010

model 1			model 2		
Fixed terms	<i>F</i> -value		Fixed terms	<i>F</i> -value	
MONTH	F(11,249)=13.69	***	MONTH	F(11, 249)=13.59	***
log(div)	F(1,21)=6.86	*	lin(age)	F(1,21)=0.26	n.s.
lin(age)	F(1,21)=2.90	n.s.	AGE	F(1,21)=1.37	n.s.
AGE	F(1,21)=0.075	n.s.	log(div)	F(1,21)=7.72	*
MONTH×log(div)	F(11,248)=0.51	n.s.	MONTH×lin(age)	F(11, 248)=0.62	n.s.
MONTH×lin(age)	F(11,249)=0.88	n.s.	MONTH×AGE	F(11, 250)=0.81	n.s.
MONTH×AGE	F(11,250)=0.89	n.s.	MONTH×log(div)	F(11,249)=0.84	n.s.
log(div)×lin(age)	F(1,21)=0.78	n.s.	lin(age)×log(div)	F(1,21)=0.78	n.s.
Log(div) ×AGE	F(1,21)=0.06	n.s.	AGE×log(div)	F(1,21)=0.06	n.s.
Random terms	Variance component		Standard error of variance component		
plot	23.83		8.137		
plot × MONTH	21.44		2.707		
residuals	41.93		1.671		

415 F values and corresponding degrees of freedom (numerator and denominator d.f.) are
416 given in brackets. YEAR (6 levels), AGE (3 levels) MONTH (12 levels) are fixed-
417 effects factors, while div, log(div) and lin(age) are continuous integer variables, i.e.
418 lin(age) the linear contrasts of the factor AGE. The fixed-effect term div stands for
419 canopy tree species richness, log(div) for the logarithm of div. Fixed-effects terms in
420 bold face indicate significance: *** $P < 0.001$, ** $P < 0.01$, * $P < 0.05$. Marginal
421 significance: · $P < 0.1$, not significant: n.s.).

422

423 **Figure legends**

424 Figure 1: total litterfall per year in different successional ages (see legend inside
425 figure). Thick solid line indicates the linear regression line based on the 27 plot means
426 pooled across successional ages. The grey shadow shows the 95% confidence
427 interval.

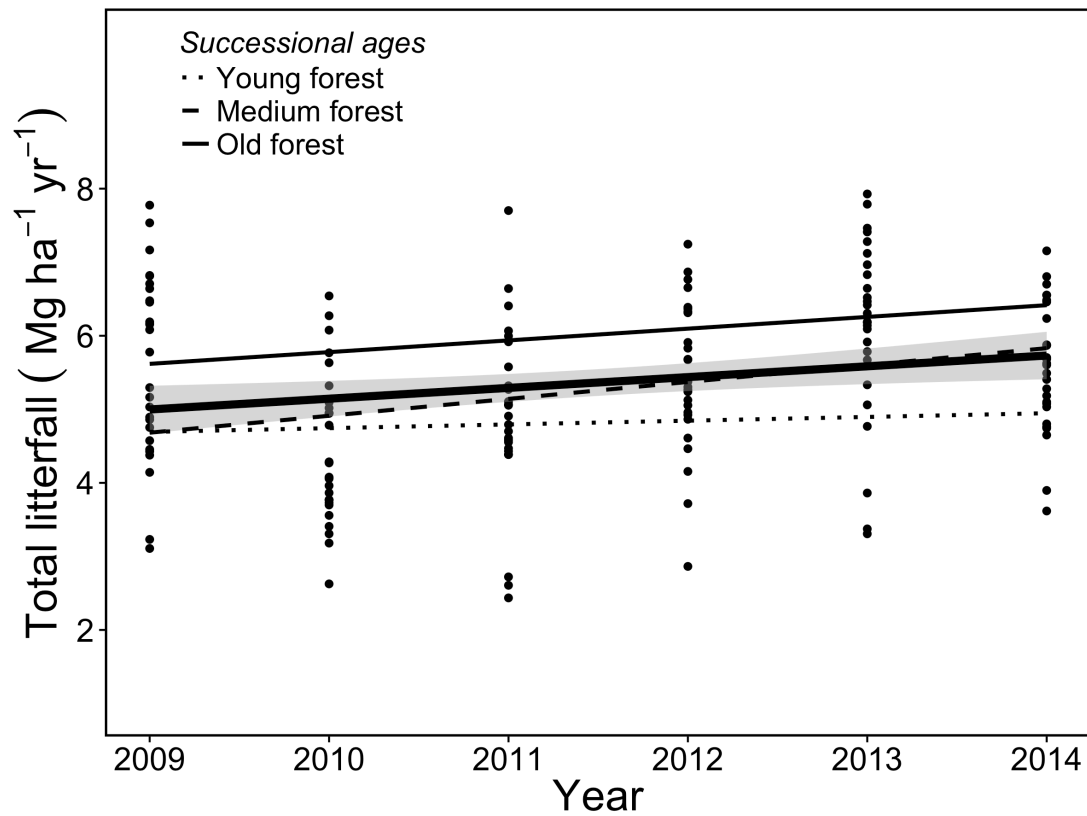
428 Figure 2: (a) total litterfall per year and (b) leaf litter C/N (mass ratio) as functions of
429 species richness in different successional ages. Error bars indicate means \pm standard
430 errors (n = 6 for total litterfall, n = 12 for C/N).

431 Figure 3: (a) total monthly litter production, (b) monthly leaf litter production and (c)
432 monthly non-leaf litter production at different species richness levels. Error bars
433 indicate means \pm standard errors (n = 9). Circles with solid line refer to high-diversity
434 plots; squares with dashed line refer to medium-diversity plots; triangles with dotted
435 line refer to low-diversity plots.

436 Figure 4: monthly leaf C/N (mass ratio) at different species richness levels. Error bars
437 indicate means \pm standard errors (n = 9). Circles with solid line refer to high-diversity
438 plots; squares with dashed line refer to medium-diversity plots; triangles with dotted
439 line refer to low-diversity plots.

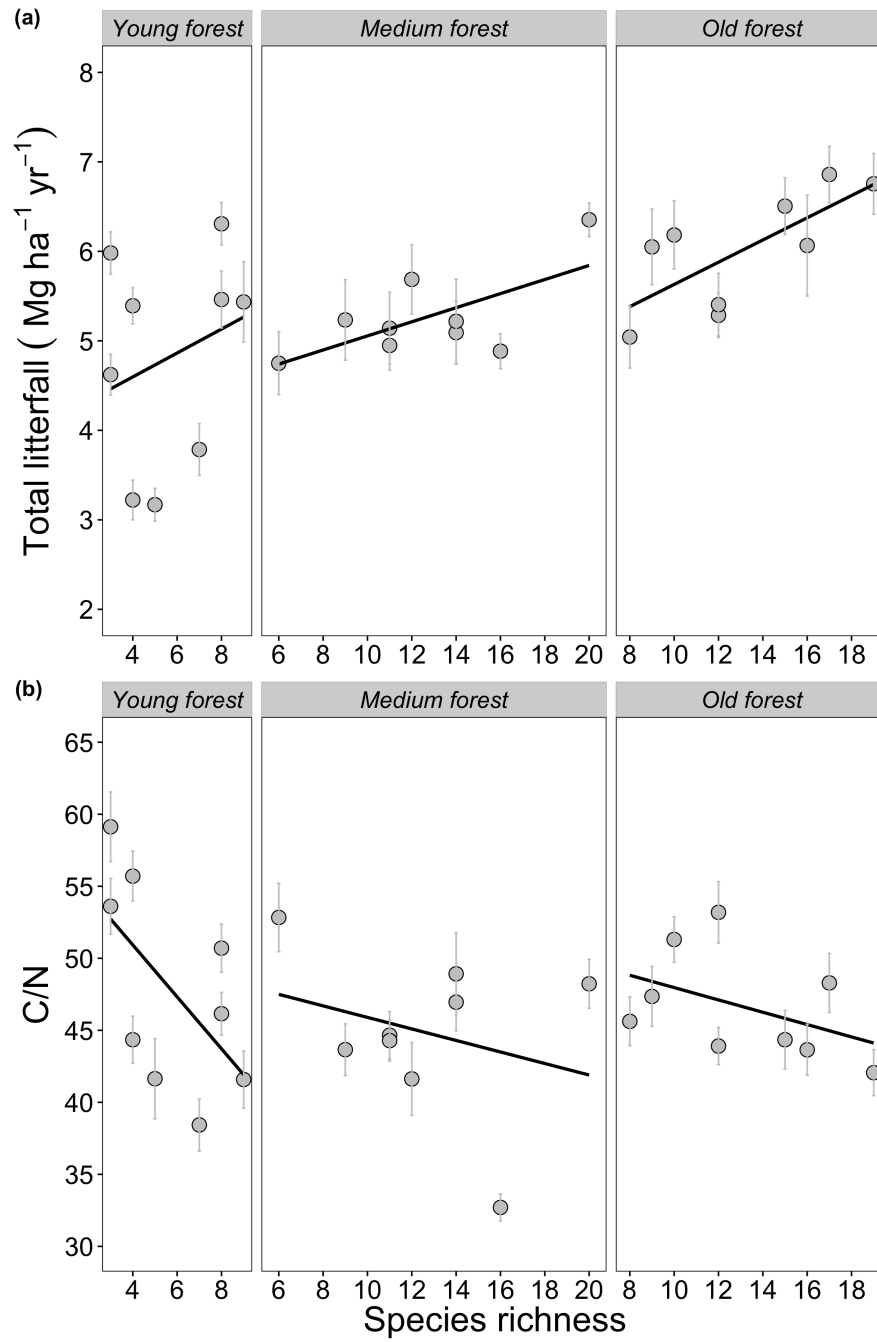
440 Figure 5: monthly litter species number per trap at different species richness levels for
441 deciduous and evergreen species and for all species combined. Error bars indicate
442 means \pm standard errors (n = 9). Circles refer to high-, squares to medium- and
443 triangles to low-diversity plots.

444 Figure 6: production of different species' leaf litter at (a) different species richness
445 levels and (b) along different successional stages (see legends inside figure). Error
446 bars indicate means \pm standard errors (n = 9).



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3 Figure 1: total litterfall per year in different successional ages (see legend inside
4 figure). Thick solid line indicates the linear regression line based on the grand mean
5 (n=27). The grey shadow shows the 95% confidence interval.



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8 Figure 2: (a) total litterfall per year and (b) leaf litter C/N (mass ratio) as functions of
9 species richness in different successional ages. Error bars indicate means \pm standard
10 errors (n = 6 for total litterfall, n = 12 for C/N).

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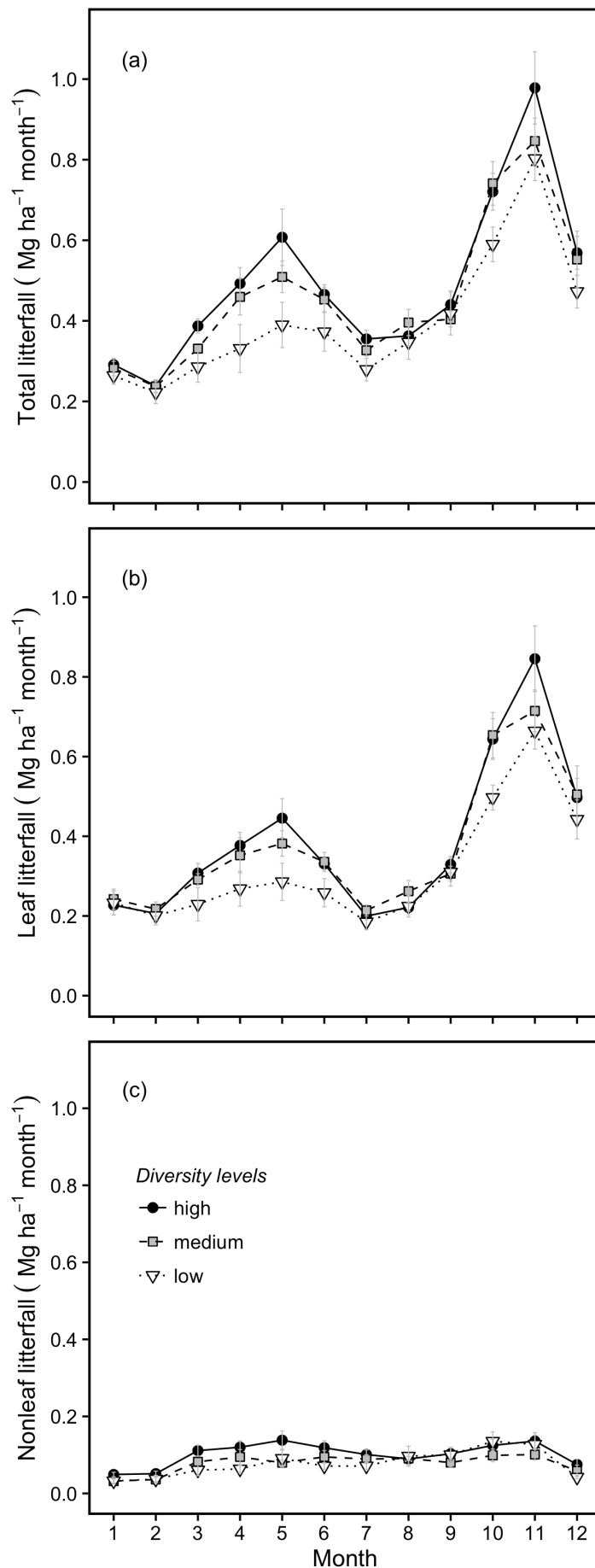
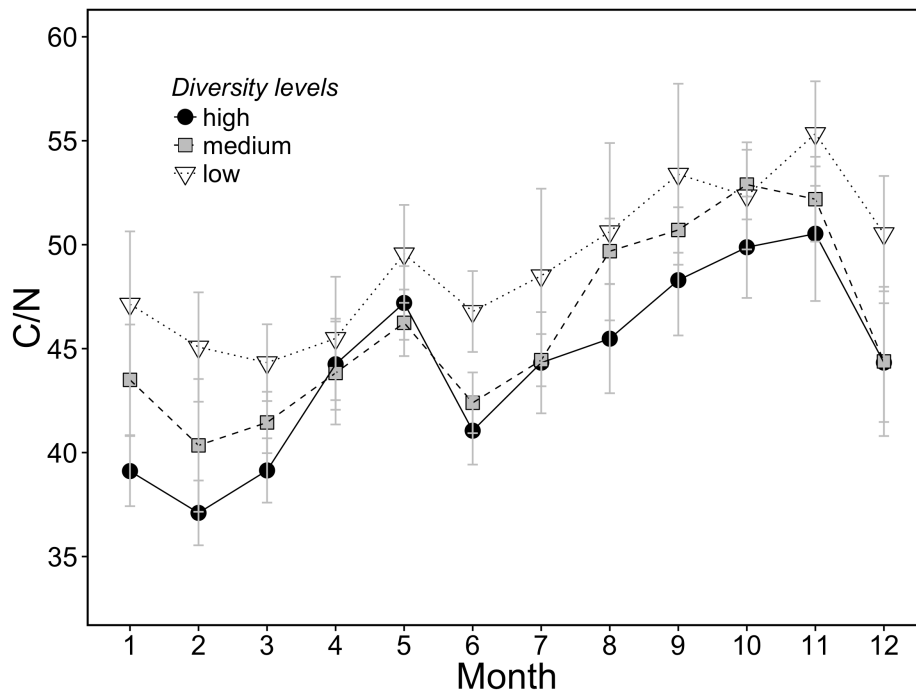
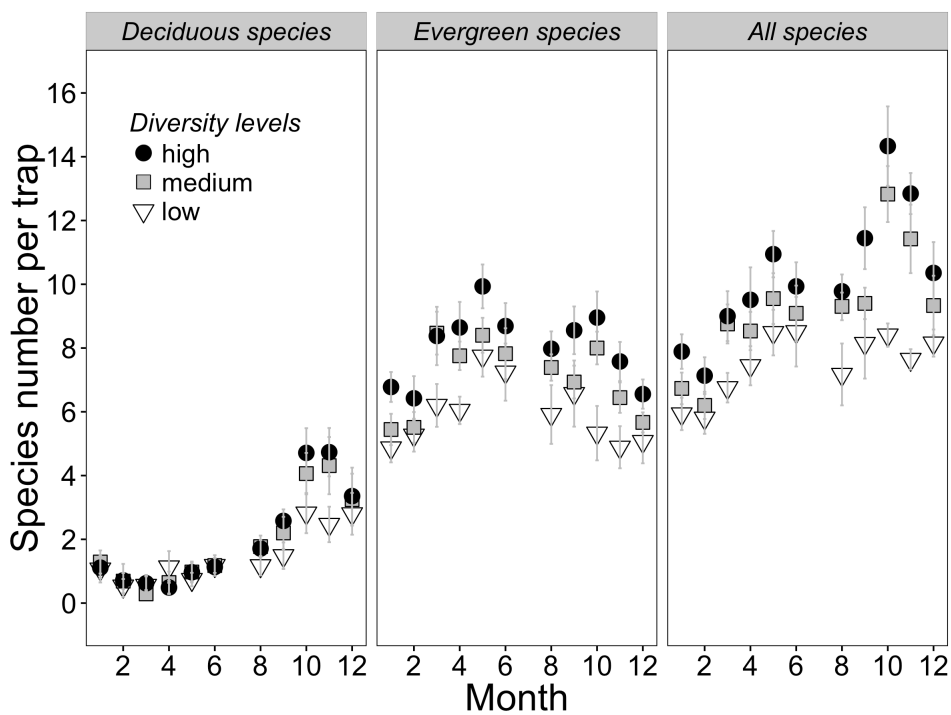


Figure 3: (a) total monthly litter production, (b) monthly leaf litter production and (c) monthly non-leaf litter production at different species richness levels. Error bars indicate means \pm standard errors ($n = 9$). Circles with solid line refer to high-diversity plots; squares with dashed line refer to medium-diversity plots; triangles with dotted line refer to low-diversity plots.



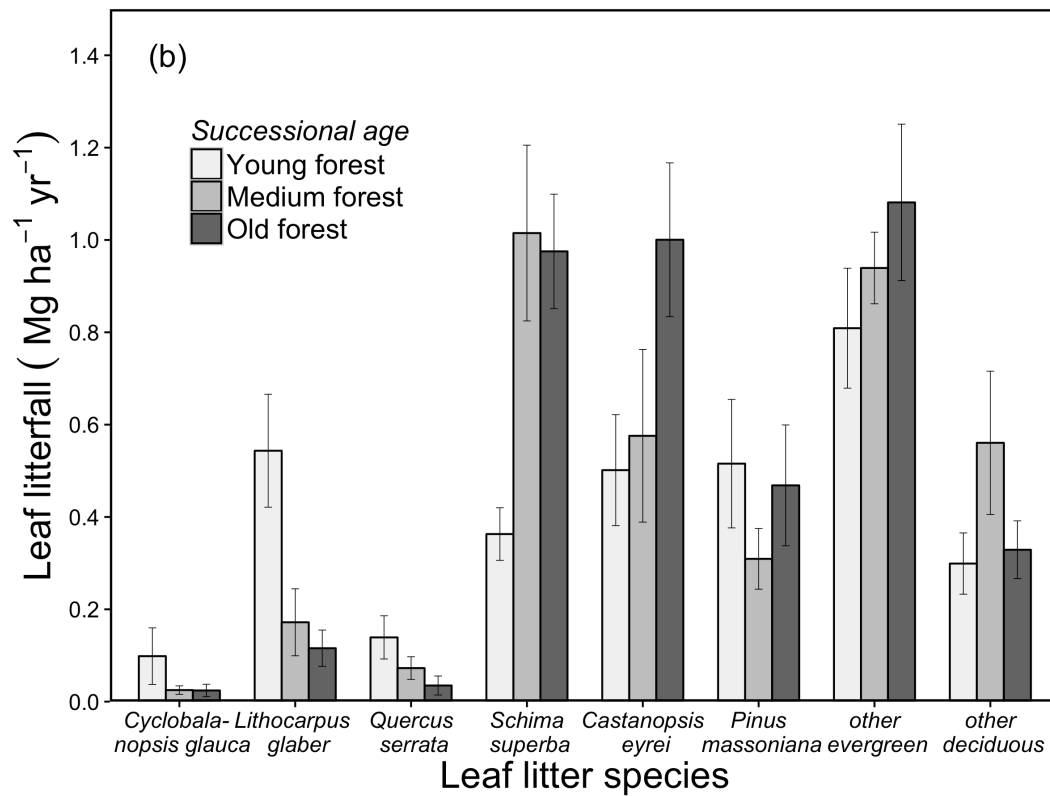
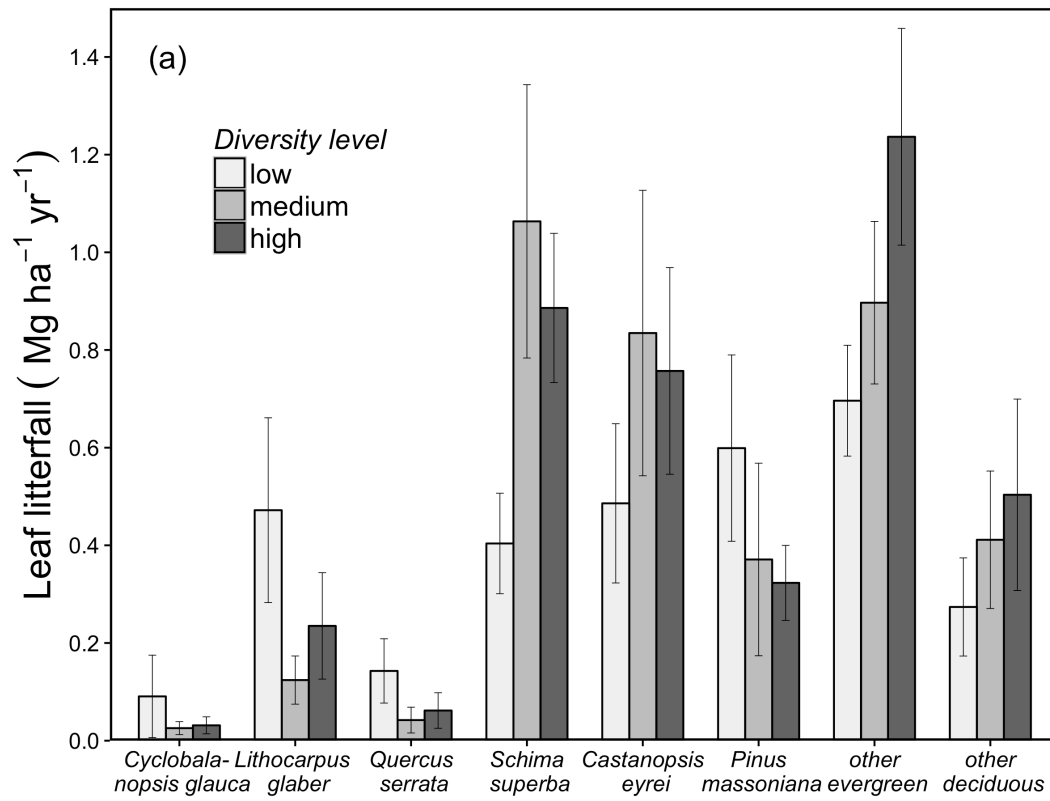
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38 Figure 4: monthly leaf C/N (mass ratio) at different species richness levels. Error bars
39 indicate means \pm standard errors ($n = 9$). Circles with solid line refer to high-diversity
40 plots; squares with dashed line refer to medium-diversity plots; triangles with dotted
41 line refer to low-diversity plots.



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44 Figure 5: monthly litter species number per trap at different species richness levels for
45 deciduous and evergreen species and for all species combined. Error bars indicate
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